Accepted Manuscript

Development of non-linear models predicting daily fine particle concentrations using aerosol optical depth retrievals and ground-based measurements at a municipality in the Brazilian Amazon region

Karen dos Santos Gonçalves, Mirko S. Winkler, Paulo Roberto Benchimol Barbosa, Kees de Hoogh, Paulo Eduardo Artaxo Netto, Sandra de Souza Hacon, Christian Schindler, Nino Künzli

PII: S1352-2310(18)30219-X

DOI: 10.1016/j.atmosenv.2018.03.057

Reference: AEA 15927

To appear in: Atmospheric Environment

Received Date: 19 June 2017

Revised Date: 19 March 2018

Accepted Date: 28 March 2018

Please cite this article as: Gonçalves, K.d.S., Winkler, M.S., Benchimol Barbosa, P.R., de Hoogh, K., Artaxo Netto, P.E., de Souza Hacon, S., Schindler, C., Künzli, N., Development of non-linear models predicting daily fine particle concentrations using aerosol optical depth retrievals and ground-based measurements at a municipality in the Brazilian Amazon region, *Atmospheric Environment* (2018), doi: 10.1016/j.atmosenv.2018.03.057.

This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.



1	Development of non- linear models
2	predicting daily fine particle concentrations
3	using aerosol optical depth retrievals and
4	ground-based measurements at a
5	municipality in the Brazilian Amazon region
6	Karen dos Santos Gonçalves*, ^{†,‡,§} , Mirko S Winkler ^{†,‡} , Paulo Roberto Benchimol
7	Barbosa ^{II} , Kees de Hoogh ^{†,‡} , Paulo Eduardo Artaxo Netto [⊥] , Sandra de Souza Hacon [§] ,
8	Christian Schindler ^{†,‡} ,, Nino Künzli ^{†,‡}
9	
10	AUTHORS ADRESS:
11	[†] Swiss Tropical and Public Health Institute, Basel, Switzerland;
12	[*] University of Basel, Basel, Switzerland;
13	§ National School of Public Health Sergio Arouca, Oswaldo Cruz Foundation –
14	ENSP/FIOCRUZ, Rio de Janeiro, Brazil;
15	^I Clinical Coordination of Pedro Ernesto University Hospital, Rio de Janeiro State
16	University – HUPE/UERJ, Rio de Janeiro, Brazil;
17	¹ Physics Institute, University of São Paulo – IFUSP/USP, São Paulo, Brazil
18	
19	
20	KEYWORDS: Aerosol Optical Depth; Particulate matter; Air pollution; Forest fire;
21	vandation approach, drazman Amazon Region.

23 ABSTRACT

24

25 Epidemiological studies generally use particulate matter measurements with diameter 26 less 2.5µm (PM_{2.5}) from monitoring networks. Satellite aerosol optical depth (AOD) 27 data has considerable potential in predicting PM_{2.5} concentrations, and thus provides an 28 alternative method for producing knowledge regarding the level of pollution and its 29 health impact in areas where no ground PM_{2.5} measurements are available. This is the 30 case in the Brazilian Amazon rainforest region where forest fires are frequent sources of 31 high pollution. In this study, we applied a non-linear model for predicting PM_{2.5} 32 concentration from AOD retrievals using interaction terms between average 33 temperature, relative humidity, sine, cosine of date in a period of 365,25 days and the square of the lagged relative residual. Regression performance statistics were tested 34 35 comparing the goodness of fit and R^2 based on results from linear regression and nonlinear regression for six different models. The regression results for non-linear 36 37 prediction showed the best performance, explaining on average 82% of the daily PM_{2.5} concentrations when considering the whole period studied. In the context of Amazonia, 38 39 it was the first study predicting PM_{2.5} concentrations using the latest high-resolution AOD products also in combination with the testing of a non-linear model performance. 40 41 Our results permitted a reliable prediction considering the AOD-PM_{2.5} relationship and 42 set the basis for further investigations on air pollution impacts in the complex context of 43 Brazilian Amazon Region.

44 45 46

1. INTRODUCTION

In spite of the efforts to improve air quality during the past decades, levels of air pollution experienced by human populations continue to cause a large burden of disease.^{1,2,3} Atmospheric aerosols and particulate matter that are breathable (< 2.5 μ m diameter = PM_{2.5}) and inhalable (< 10 μ m = PM₁₀), generated from natural and anthropogenic emission sources present known effects for a number of causes of death, particularly the increase in cardio-respiratory diseases in areas with high concentrations.^{4,5}

Intensive and indiscriminate occurrence of forest fire has become a serious environmental problem in Brazil, affecting ecosystems' balance and human health with consequences at the local, regional and global level.^{6,7} Brazilian Amazon region has geographic and environmental circumstances that are distinct from other world regions.

58 For this reason, the occurrence of fire and emissions of $PM_{2.5}$ exposes every year 59 increasingly large portions of vulnerable populations.^{8,9}

To understand the association between $PM_{2.5}$ and effects on human health, epidemiological studies have employed $PM_{2.5}$ measurements from monitoring sites. However, due to cost and lack of appropriate infrastructure, especially in rural and remote areas, no fixed site $PM_{2.5}$ measurements are available in many regions of Brazil. This is a major limitation for estimating exposure to $PM_{2.5}$ and assessing health impacts associated with forest fires as one of its major source.^{10,11,12,13,14,15}

An alternative approach to estimate the air quality in areas without direct PM_{2.5} 66 measurements is by means of satellite remote sensing using aerosols optical depth 67 (AOD). AOD is an electromagnetic radiation measure and reflects the integrated 68 number of particles at a given wavelength. It is an important satellite-retrieved property 69 for predicting the PM_{2.5} concentrations due repeated observations of the atmosphere and 70 its extensive spatial coverage.¹⁶ The AOD has been successfully used in statistical 71 models for estimating PM_{25} levels. As shown by previous studies, parameters such as 72 73 local meteorology and land use information influence the relationship between AOD 74 and daily PM_{2.5} concentrations, which need to be considered as additional predictors.^{10,17,18,19,20,21,22,23,24} 75

Traditionally, the health exposure studies have used the standard MODIS (Moderate
Resolution Imaging Spectroradiometer) AOD product of the "Dark Target" algorithm
published by Levy et al. (2007, 2010), which has a resolution of 10 x 10 km². Later,
Remer et al. (2013, 2005) described AOD algorithm applying a higher resolution of 3 X
3 Km².^{25,26,27,28}

Concerning the applicability of the statistical methods for predicting PM_{25} 81 concentration using AOD retrievals, de Hoogh K et al $(2017)^{29}$ used a higher spatial 82 83 resolution for modelling daily PM_{2.5} concentrations across Switzerland during the period between 2003 to 2013. Their models result explained on average 73% of the total 84 ,71% of the spatial and 75% of the temporal variation (all cross validated) in measured 85 PM_{2.5} concentrations. Kloog Itai et al. (2017)³⁰ described a new hybrid spatio-temporal 86 model for estimating daily PM_{2.5} concentrations across northeastern USA using high 87 88 resolution AOD data. Their results showed a high predictive accuracy at high spatial 89 resolutions using a mixed model regressing PM_{2.5} measurements with an excellent 90 model performance (R2=0.88).

3

92 These recent studies still have the challenge of reducing exposure error, although 93 shows better fits than previous models. In spite our model showed a good performance, 94 it is important to reproduce it in another region with di erent meteorological and 95 geographical patterns. Our model can be applied to other sites if site-specific AOD and 96 meteorological data are available, to be inserted into the prediction equation. The lagged 97 relative residual added as a further predictor variable it is cautious strategy to remove 98 the serial autocorrelation and to further improve the model. As another important 99 challenge is that AOD data availability is much greater in the dry seasons compared to the rainy period. This is mostly due to heavily clouded days which results in missing 100 101 AOD data. This non-random lack of AOD readings could negatively a lect predictive 102 performance. Also, treating large areas, such as Brazilian Amazon region, can add 103 additional selection bias since there may be meteorological variations in the daily calibration between PM2.5 and AOD.³⁰ 104

In this paper we developed a non-linear model predicting daily fine particle concentrations using AOD retrievals at 3 x 3 km resolution and ground-based measurements at a municipality of Porto Velho, Brazil during the period between 2009 to 2011. For Brazilian Amazon region, it is the first study to develop this approach considering a non-linear model predicting PM2.5 concentrations. This study assessment is part of an investigation that aims at analysing the impact of PM2.5 exposure on cardiovascular disease in Porto Velho.

- 112
- 113

2. MATERIAL AND METHODS

114 **2.1. Ground-level PM_{2.5} data**

115 Daily averages were derived during the period from 25 September 2009 to 21 116 October 2011 with a total of 757 days. Over the study period, PM_{2.5} concentrations were 117 measured for 24h at one air quality monitoring station in Porto Velho municipality, 118 which was implanted in partnership between Institute of Physics at University of São 119 Paulo (USP), University of Rondônia (UNIR), Environmental Biogeochemistry 120 Laboratory Wolfgang H. Pfeiffer and National School of Public Health-Oswaldo Cruz 121 Foundation (FIOCRUZ) in Brazil. The PM_{2.5} monitor is located at 15 km north of to the 122 centre of urban area (Figure 1). Porto Velho municipality is the third capital in the 123 Brazilian Amazon region with 67 districts within the urban area. With an area of 34,096 km² Porto Velho has a population of 503,000 inhabitants according to Brazilian 124 125 Institute of Geography and Statistics (IBGE, Census 2010). PM_{2.5} measurements were

collected by means of a stacked filter unit (SFU) and were analysed gravimetrically
according to the World Health Organization Air Quality Guidelines for particulate
matter, ozone, nitrogen dioxide and sulfur dioxide (WHO, 2005)³¹.

This methodology involves the sampling site (8.69° S, 63.87° W) located in a region with large land use changes and associated regional biomass burning. The SFU (Stacked Filter Unit) type samplers and the analysis follows routine gravimetric techniques³³. In addition, trace elements and ionic compounds are collected, allowing for future analyses.

134 There is an AFG sampler, which collects aerosols for elemental PIXE and black 135 carbon analyses on the roof of the shelter, 24 hours sampling. The collection of aerosol 136 particles using filters is a simple and very common method for sampling aerosol 137 particles. Filters allow elemental and ionic analysis through a series of measurement 138 techniques. The sampler collects fine and coarse particles and contains an inlet that 139 allows the entry of particles in the range of $2 < Dp < 10\mu m$. The filters are polycarbonate, 140 having a diameter of 47mm and are arranged in series. In the first step the particles of 141 the coarse fraction are retained using Nucleopore filters with pores of 8µm in diameter, 142 in the second stage, they are the fine particles that are retained using the filter 143 Nucleopore with pores of 0.4µm. The samples collected with the AFG sampler was 144 used to determine the mass of the aerosols by means of gravimetric analysis, the concentration of black carbon and to quantify the elemental concentration of the 145 146 material deposited in the filters.

147

148 **2.2.MODIS 3 km AOD retrieval**

149 The Moderate Resolution Imaging Spectroradiometer (MODIS) is a key instrument 150 aboard the Terra and Aqua satellites of the National Aeronautics and Space 151 Administration (NASA) and has been in operation since 1999 and 2002, respectively. While Terra passes the equator in the morning, from north to south, Aqua passes the 152 153 equator from south to north in the afternoon. These satellites were used to retrieve AOD 154 aerosol products with a 3 km resolution (MOD04_3K and MYD04_3K), operating at an 155 altitude of approximately 700 km (http://modis-atmos.gsfc.nasa.gov/). In the Collection 156 6, Level 2 aerosol products, the most recent 3 km AOD dark target retrieval algorithm is 157 similar to the 10 km standard product (Collection 5, Level 2) and has three different 158 wavelength channels of 0.47, 0.66 and 2.12 µm employed for AOD retrieval over land. 159 The other channels are used for screening procedures (e.g., coverage of cloud, snow and

ice).^{25,26,32} More details on the retrieval of MODIS satellite aerosol data have previously
been published by Remer et al. (2013, 2005)^{25,28} and Levy et al. (2007, 2010)^{26,27}. For
the AOD daily averages, we used the algorithm retrieval in MATLAB (version 2015a,
MathWorks) and the software ArcGIS (version 10, ESRI) to create 820 grid cells of 3 x
3-km covering the study area for spatial analyses.

165

166 **2.3. Statistical model and validation**

In this study we considered five different types of prediction models of PM_{2.5} 167 168 concentrations from AOD retrievals. They were all of the form 169 $PM_{2.5} = exp($ linear predictor), with the linear predictor involving terms composed of AOD and other influencing factors. The advantage of such a model over one for log-170 171 transformed outcome data is that it provides estimates of mean exposure levels while estimates of geometric mean levels are obtained when modelling log-transformed 172 173 outcome data and then exponentiating the resulting predictions (which are on the log-174 scale). In Model 1, we took the linear predictor to be a cubic polynomial in AOD with 175 time-independent coefficients:

- 176
- 177
- 178

179 In a next step, we considered the coefficients of Model 1 to be polynomials of second 180 degree in average temperature (TEMP) and relative humidity (RH). Thus, each of the 181 coefficients was assumed to be of the form $\gamma_0 + \gamma_1 \cdot \text{TEMP} + \gamma_2 \cdot \text{RH} + \gamma_3 \cdot \text{TEMP}^2 + \gamma_4 \cdot \text{RH}^2$ 182 $+ \gamma_5 \cdot \text{TEMP} \cdot \text{RH}$. Multiplying these polynomials with AOD, AOD² and AOD³ and the 183 intercept, respectively, provided:

184

(Model 2)

(Model 1)

 $PM_{2.5} = \exp(\alpha + \beta_1(AOD) + \beta_2(AOD)^2 + \beta_3(AOD)^3)$

- 185 $PM_{2.5} = Model \ 1 \cdot exp(\beta_4(AOD \cdot TEMP) + \beta_5(AOD \cdot TEMP^2) + \beta_6(AOD \cdot RH) + \beta_6(AO$
- $186 \qquad \beta_7(AOD \cdot (RH^2) + \beta_8(AOD \cdot TEMP \cdot RH) + \beta_9(AOD^2 \cdot TEMP) + \beta_{10}(AOD^2 \cdot TEMP^2) + \beta$
- 187 $\beta_{11}(AOD^2 \cdot RH) + \beta_{12}(AOD^2 \cdot RH^2) + \beta_{13}(AOD^2 \cdot TEMP * RH) + \beta_{14}(AOD^3 \cdot TEMP) + \beta_{14}(AOD^3$
- $188 \qquad \qquad \beta_{15}(AOD^3 \cdot TEMP^2) + \beta_{16}(AOD^3 \cdot RH) + \beta_{17}(AOD^3 \cdot RH^2) + \beta_{18}(AOD^3 \cdot TEMP \cdot RH) \)$
- 189

In an attempt to further improve the model, we added interaction terms between AOD, AOD^2 and AOD^3 with rainy season (Model 3) and tested interactions of the three AOD-terms with sine and cosine functions of date with a period of 365.25 days in the Model 4.

(Final model)

In the last step, we included the lagged relative residual and its square as additional predictor variables (Model 5). This was to reduce serial autocorrelation. The final model obtained after some backward elimination steps was of the form:

197

- 198 $PM_{2.5} = exp(\alpha + \beta_1 \cdot AOD + \beta_2 \cdot AOD^2) + \beta_3 \cdot AOD^3 + \beta_4 \cdot (AOD \cdot TEMP) + \beta_5(AOD \cdot RH)$
- 199 + $\beta_6(AOD \cdot TEMP^2) + \beta_7(AOD \cdot RH^2) + \beta_8(AOD \cdot \cos_days) + \beta_9(AOD \cdot \sin_days) +$
 - $\beta_{10}(AOD^2 \cdot \cos_d ays) + \beta_{11}(AOD^2 \cdot \sin_d ays) + \beta_{12}(AOD^3 \cdot \cos_d ays) + \beta_{10}(AOD^3 \cdot \cos$

201
$$\beta_{13}(AOD^3 \cdot sin_days) + \beta_{14}(residual) + \beta_{15}(residual)^2$$

202

200

In these equations, $PM_{2.5}$ denotes the predicted concentrations, where exp is the 203 204 exponential function; cos_days and sin_days denote the cosine and sine terms of date 205 with a period of 365,25 days; residual denotes the lagged relative residual. Relative residuals were defined as ratio between residuals and predicted values. Model 206 207 performance was evaluated by comparing the predictions with the ground measurements using the adjusted coefficient of determination (R^2 adj), residual standard deviation 208 209 (RMSE), Akaike's information criterion (AIC), and partial autocorrelation of residuals 210 by lags. High values of adjusted R squared suggest that MODIS AOD data can be used 211 to estimate ambient concentrations. Furthermore, we calculated mean, standard 212 deviation and maximum / minimum values to summarize the descriptive statistics of our 213 sample for the whole period, the dry season (months June to November, when forest fires occur in Brazilian Amazon region) and the rainy season (months from December 214 215 to May).

Daily meteorological data on average temperature and relative humidity were obtained for the period 25 September 2009 to 21 October 2011 from the monitoring station of INMET (*National Meteorology Institute*) in Porto Velho. The information is publicly available on the website of the institute (www.inmet.gov.br).

The spatial distribution of the 3x3km resolution MODIS AOD average during the study period was derived by spatial interpolation using the inverse distance weighting (IDW). We present the results for all states in Brazil and for our study area. It is important to highlight that all regression results were presented with the original AOD and PM_{2.5} datasets. The software R (version 3.1.3) was used for statistical analyses.

225

226 **2. RESULTS**

3.1. Descriptive statistics

Table 1 shows the descriptive statistics of daily measured $PM_{2.5}$, values of AOD, relative humidity, average temperature and precipitation from 25 September 2009 to 21 October 2011, as well as for the dry and rainy seasons. Average daily level of $PM_{2.5}$ from the ground-level monitor was $11\mu g/m^3$ with a standard deviation (SD) of $\pm 20\mu g/m^3$ over the three years studied. Of note, the analysis of the data for 2010, the year when one of the most extreme dry seasons in Brazilian Amazon region occurred, revealed an annual-mean of $36\mu g/m^3$ ($\pm 46\mu g/m^3$ SD).

235 Considering the differences between seasons, the maximum daily value was 236 exceptionally high $(164\mu g/m^3)$ during the dry seasons of the period studied compared to 237 $27\mu g/m^3$ in the rainy seasons.

Over the entire study period, the daily AOD values observed varied from 0.03 to 239 2.19. On average, 649 AOD values were retrieved per grid cell which corresponds to 240 86% of the entire study period of 757 days.

All the meteorological variables such as relative humidity, average temperature and precipitation were consistent with the climatic patterns expected for the Brazilian Amazon region and thus support the analysis in the regression models.

244

245 **3.2. Non-linear prediction models**

To test the performance of the five regressions models we use a total of 649 valid days for the model fitting. The comparisons between the models analysed and parameters estimated are shown in **Table 2 and 3**.

Model 1 shows an adj R² of 0.54, RMSE of 13.59µg/m³ and AIC of 5234.3 for the 249 whole period. Model 2 including interactions between AOD, AOD^2 and AOD^3 and 250 linear and quadratic terms in temperature and relative humidity provided a better fit (R^2) 251 252 = 0.67). After adding interactions between the three AOD-terms and rain the fit only slightly improved ($R^2=0.70$). In Model 4 we excluded the rain term and added 253 254 interactions of the three AOD-terms with sine and cosine of date with a period of 365.25 days. This model performed considerably better ($R^2=0.77$; RMSE=9.59µg/m³; 255 256 AIC=4803.1).

After adding the lagged relative residual and its square as additional predictors (Model 5) the adjusted R2 further increased to 0.82 (RMSE= $8.60\mu g/m^3$; AIC=4797.6). This means that this non-linear prediction model explains 82% of the variance of daily PM_{2.5} concentrations in combination with meteorological and seasonal variables. The introduction of these two terms also led to a drastic reduction in residual

autocorrelation, in that lag1-autocorrelation of residuals was no longer significant 262 263 (Figure 2). As visualized in Figure 2, the time series of predicted $PM_{2.5}$ concentrations 264 follows a very similar pattern as the measured PM_{2.5} confirming the high performance 265 of the prediction models. The period from mid-July to end of Octobre 2010 – a dry 266 period with plumes of biomass burning – is characterized by very high AOD values, 267 reaching peaks 50-100 times above the typical values observed before and after this 268 period. The comparisons between the measured and predicted PM_{2.5} concentrations for 269 Model 1 and Model 5 are illustrated in Figure 3.

- 270 The Spatial distribution of PM_{2.5} predicted concentration over the basin during 271 different seasons for all Brazilian states are shown in Figure 4. The highest predicted 272 PM_{2.5} concentrations were observed in the Brazilian Amazon region during the forest 273 fires season (months between September, October and November). In our study area, 274 $PM_{2.5}$ averages reached $44\mu g/m^3$ in the urban area of Porto Velho, and $54 \mu g/m^3$ across 275 the Rondônia state during the forest fires between 2009 and 2011. Information about the 276 distribution of PM_{2.5} within the urban districts of Porto Velho and the relation with the 277 health data will be presented in a separate manuscript about the impacts of PM_{2.5} on 278 human health in Brazilian Amazon region.
- 279

280 **3. DISCUSSION**

281 The results of our final non-linear prediction model for PM_{2.5} showed a good 282 performance, explaining on average 82% of the variance in measured PM_{2.5} concentrations during the period studied. This result is similar and in accordance with 283 the findings presented by Lee et al $(2011)^{10}$ and Xie et al $(2015)^{24}$, who showed 284 285 prediction models that explained 92% and 82% of the variance in PM_{2.5} concentrations 286 in the North-eastern, US and in Beijing, China, respectively. Our model has the 287 advantage that it does not produce negative predictions and fits the mean of the data as a 288 function of the predictor variables. Moreover, by including the lagged relative residual 289 and its square as additional predictor variables it was possible to remove the significant 290 lag1-autocorrelation, and to further improve the model fit.

Observing the temporal distribution of predicted and measured $PM_{2.5}$ concentrations it is important to highlight the enormous peak of $PM_{2.5}$ observed between days 300 and 400 during the dry periods in our study area with the maximum daily value of 164µg/m³. This value more than 6.5 times higher than the daily mean guideline value proposed by WHO to protect public health (25µg/m³). During the dry season, this value

296 was exceeded on X% of all days. As a consequence, the long-term mean concentration 297 during the dry period was 2 times above the WHO annual mean guideline value, set at 298 10µg/m³. Currently, these values were adopted in only a few countries as legally 299 binding targets, thus, policy makers accept major impacts on morbidity and mortality.³⁴ On the other hand, during the rainy seasons concentrations were low $(2\mu g/m^3; \pm 3\mu g/m^3)$ 300 301 SD) and fully in line with both, the daily and annual targets proposed by WHO. This 302 confirms the dominant role of fires as source of ambient air pollution in the Amazon 303 region. This result highlight the importance to set limits for PM_{2.5} in the Brazil Air 304 quality Standards defined by the National Environmental Agency (CONAMA) that currently set limits only for PM₁₀.³⁴ Our data provide unique input to evaluate whether 305 306 the CONAMA may add to the Brazil Air Q S limits for PM_{2.5} rather than PM₁₀ alone. 307 This will be particularly worth if sources and spatio-temporal patterns of the two 308 markers of air pollution largely vary across Brazil.

The non-linear prediction model demonstrated a high performance in predicting the 309 daily PM_{2.5} concentrations. However, some limitations, such as cloud properties and 310 311 uncertainties need to be mentioned. The use of only one air monitoring station for the 312 development and evaluation of the model is a major limitation of this study. However, 313 although this limits our ability to draw firm conclusions about the applicability of the 314 model across Brazil, it does provide a valid approach to predict PM_{2.5} across our main 315 study area. PM_{2.5} is spatially rather homogenously distributed, thus extrapolation of the model from the measurement site to the adjacent urban area of Porto Velho is expected 316 317 to be reliable. The model gives also good indications of possible hot spots of pollution 318 where it may be worth installing additional monitors operating continuously or at least 319 during dry seasons. With the use of mobile stations, one could characterize the spatial 320 pattern of air pollution across a larger area with only one or a few monitors while using 321 the current central monitoring station as a reference point to understand the temporal 322 variation. The installation of air quality monitoring networks is an important step for the 323 future evaluation of progress in clean air management and the assessment of its health 324 impact.

The predictions were based on different spatial scales which may be a source of uncertainties. The AOD satellite data is based on a grid cell with 3km resolution while $PM_{2.5}$ ground-level is measured at a fixed point. It is also important to highlight the complex relationship between AOD and $PM_{2.5}$ due the nature of the forest aerosol in

Brazilian Amazon. Other important predicting factors such as wind speed, atmosphereor physical-chemical components were not analysed in this study.

331 Despite the uncertainties, this study is the first to predict $PM_{2.5}$ concentrations using 332 a non-linear prediction model and the higher-resolution MODIS AOD products in Porto 333 Velho. By modelling the AOD- $PM_{2.5}$ relationship in a time-dependent manner reflecting 334 seasonal fluctuations and influences of temperature and relative humidity we were able 335 to develop a prediction model for PM25 with good fitting properties.

- 336 Our model also can be applied to other sites of the region if site-specific AOD- and meteorological data are available, to be inserted into the prediction equation. On the 337 338 other hand, as measured PM_{2.5} data are only available for the reference station, the 339 relative lag1-residuals of PM_{2.5} at the reference site would have to be used for all other 340 sites. About the applicability of the model to other sites, the Model 5 (not involving 341 lagged residuals) could be applied to any other site of the Amazon region, as AOD-342 values and estimates of meteorological parameters can be obtained for any such site 343 based on satellite data. On the other hand, model 6 involves an input variable which is 344 only available at the reference site and whose values would therefore have to be used at all other sites. A cautious strategy might therefore consist in using both model 5 and 6 345 346 when conducting time series analyses of deaths and hospital admissions.
- 347

348 **4. CONCLUSIONS**

349 Satellite data has an important potential for the spatio-temporal prediction of $PM_{2.5}$ 350 concentrations. It offers an alternative method to describe the impacts of forest fires on 351 air quality and to assess the related health effects in the Brazilian Amazon region. Our 352 method provides valuable inputs on how to strengthen and optimize the $PM_{2.5}$ 353 monitoring networks with well-placed complementary measurement sites.

This is needed to understand the impact of air pollution in Brazil and to demonstrate the improvements of air quality and the related health benefits due to the adoption of clean air policies.

357

358 AUTHOR INFORMATION

359 **Corresponding author:**

360 ***Postal address:** ENSP/FIOCRUZ: Rua Leopoldo Bulhões, 1480, CEP: 21041-210,

- 361 Manguinhos, Rio de Janeiro, Brazil; E-mail: <u>karengoncalves@gmail.com</u>
- 362

363 **Note:** The authors declare no conflict of interest.

364 365

366 367

ACKNOWLEDGEMENT

368 This work was supported by the CAPES foundation [grant number 88881.068027/2014-01] from Brazil. Special thanks to Glauber G.C. Silva 369 370 (INPE/Brazil), who was always kind and available to help with MATLAB scripting and 371 technical solutions. This manuscript is part of a larger project on *Cardiovascular* 372 diseases and the exposure to forest fires in Porto Velho municipality, Rondônia state, 373 Brazil that was submitted and accepted by the Research Ethics Committee of the Sergio 374 Arouca National Public Health School (Comitê de Ética em Pesquisa da Escola Nacional de Saúde Pública Sergio Arouca – ENSP / FIOCRUZ) according to 375 376 Resolution number 466/2012 from National Research Ethics Council (Conselho 377 Nacional de Pesquisa – CONEP) under CAAE number 41732615.4.0000.5240.

378379

380 **REFERENCES**

- (1) Cohen, AJ; Anderson, HR; Ostro, B; Pandey, KD; Krzyzanowski, M; Künzli, N;
 Gutschmidt, K; Pope, A; Romieu, I; Samet, JM & Smith, K. The Global Burden of
 Disease Due to Outdoor Air Pollution. *Journal of Toxicology and Environmental Health, Part A*, 2005. 68:1301–1307, doi:10.1080/15287390590936166
- 385

(2) Brauer M, Freedman G, Frostad J, van Donkelaar A, Martin RV, Dentener F, Van
Dingenen R, Estep K, Amini H, Apte JS, Balakrishnan K, Barregard L, Broday
DM, Feigin V, Ghosh S, Hopke PK, Knibbs LD, Kokubo Y, Liu Y, Ma S,
Morawska L, Sangrador JLT, Shaddick G, Anderson HR, Vos T, Forouzanfar MH,
Burnett RT, Cohen A. Ambient air pollution exposure estimation for the Global
Burden of Disease 2013. *Environmental Science & Technology*. 2015 Nov 23. doi:
10.1021/acs.est.5b03709.

393

394 (3) Global Burden of Diseases (GBD) (2010) Institute for Health Metrics and
 395 Evaluation 2013. <u>http://vizhub.healthdata.org/irank/heat.php</u>

- 396
- 397 (4) Brook, RD; Rajagopalan, S; Pope III, CA; Brook, JR; Bhatnagar, A; Diez-Roux,
 398 AV; Holguin, F; Hong, Y; Luepker, RV; Mittleman, MA; Peters, A; Siscovick, D;

	DDT					TDT
	ΗРΙ	N/L	ΑΝ		EK	
1000		1.1.1.1				

399	Smith, SC; Whitsel, L & Kaufman, JD. Particulate matter air pollution and
400	cardiovascular disease: An update to the scientific statement from the American
401	Heart Association. Journal AHA. Circulation 2010. 2331-2378, doi:
402	10.1161/CIR.0b013e3181dbece1;
403	
404	(5) World Health Organization (WHO) (2014) 7 million premature deaths annually
405	linked to air pollution. http://www.who.int/mediacentre/news/releases/2014/air-
406	pollution/en/
407	
408	(6) Gonçalves, KS; Castro, HA; Hacon, SS. As queimadas na região amazônica e o
409	adoecimento respiratório. Rev Ciência e Saúde Coletiva 2012. 17(6): 1523-1532,
410	doi: http://dx.doi.org/10.1590/S1413-81232012000600016;
411	
412	(7) Becker BK. Geopolítica da Amazônia. Estudos avançados 2005; 1(53):71-86.
413	
414	(8) Fearnside, P.M. Desmatamento na Amazônia brasileira: história, índices e
415	consequências. MEGADIVERSIDADE, Vol 1 (1) Julho 2005, Available:
416	http://philip.inpa.gov.br/publ_livres/2005/Desmatamento%20historia-
417	Megadiversidade.pdf, Accessed: May/2016;
418	
419	(9) Goncalves,KS; Siqueira,ASP,Castro, HA; Hacon, SS.Indicator of socio-
420	environmental vulnerability in the Western Amazon. The case of the city of Porto
421	Velho, State of Rondônia, Brazil.Ciência & Saúde Coletiva, 19(9):3809-3817, 2014
422	doi: 10.1590/1413-81232014199.14272013
423	(10) Lee H. L. Liu, Y. Coull, P.A. Schwartz, L. and Koutrakis, P. A. calibration
+24	(10) Lee, H. J., Elu, T., Coull, B.A., Schwartz, J. and Koutrakis, F. A canoration method of MODIS AOD date to predict DM2.5. Atmos. Chem. Phys. 11, 7001
+23	8002 2011 doi: 10.5104/com 11.7001 2011
+20	8002, 2011- doi: 10.3194/acp-11-7991-2011
+27	(11) Ducker D. Schneiden A. Dreitnen S. Comm. J. Deters A. Heelth offects of
428	(11) Ruckeri R, Schneider A, Breitner S, Cyrys J, Peters A. Health effects of
+29 120	particulate air pollution: A review of epidemiological evidence. <i>Inhal Toxicol</i> .
+30	2011Aug; 23(10):333-92, doi: 10.3109/08938378.2011.393387.
+31	

			a nh					10	\mathbf{C}	\cap		1.11	D.	
\mathbf{A}		·Ρ			\mathbf{N}	14	<u> </u>				к		24	
	\sim					1.17.4			$\mathbf{\nabla}$	Ú.	11/2			

- 432 (12) Ye X, Wolff R, Yu W, Vaneckova P, Pan X, Tong S. Ambient Temperature and
 433 Morbidity: A Review of Epidemiological Evidence. *Environ Health Perspect.* 2011
 434 Aug, doi: 10.1289/ehp.1003198.
- 435

436 (13) Yi O, Hong YC, Kim H. Seasonal effect of PM(10) concentrations on mortality
437 and morbidity in Seoul, Korea: a temperature-matched case-crossover analysis.
438 *Environ Res.* 2010 Jan;110(1):89-95, doi: 10.1016/j.envres.2009.09.009.

- 439
- 440 (14) Arbex MA, de Souza Conceição GM, Cendon SP, Arbex FF, Lopes AC, Moysés
 441 EP, et al. Urban air pollution and chronic obstructive pulmonary disease-related
 442 emergency department visits. *J Epidemiol Community Health*. 2009
 443 Oct;63(10):777-83, doi: 10.1136/jech.2008.078360 ;
- 445 (15) McMichael AJ, Wilkinson P, Kovats RS, Patternden S, Hajat S, Armstrong B,
 446 et al. International study of temperature, heat and urban mortality: the
 447 'ISOTHURM' project. *Int J Epidemiol.* 2008 Oct;37(5):1121-31, doi:
 448 10.1093/ije/dyn086
- 449

444

450 (16) Kloog, I.; Chudnovsky, A. A.; Just, A. C.; Nordio, F.; Koutrakis, P.; Coull, B.
451 A.; Lyapustin, A.; Wang, Y.; Schwartz, J., A new hybrid spatio-temporal model for
452 estimating daily multi-year PM 2.5 concentrations across northeastern USA using
453 high resolution aerosol optical depth data. *Atmospheric Environment* 2014, 95, 581454 590, doi:10.1016/j.atmosenv.2014.07.014

- 455
- 456 (17) Liu, Y., Sarnat, J. A., Kilaru, V., Jacob, D. J., and Koutrakis, P.: Estimating
 457 ground-level PM2.5 in the eastern United States using satellite remote sensing,
 458 Environ. Sci. Technol., 39, 3269–3278 ; 2005.
- 459

460 (18) Liu, Y., Franklin, M., Kahn, R., and Koutrakis, P.: Using aerosol optical thickness
461 to predict ground-level PM2.5 concentrations in the St. Louis area: A comparison
462 between MISR and MODIS, Remote Sens. Environ., 107, 33–44, 2007a.

- 463
- 464 (19) Liu, Y., Koutrakis, P., and Kahn, R.: Estimating fine particulate matter component
 465 concentrations and size distributions using satellite-retrieved fractional aerosol
 466 optical depth: Part 1 Method development, J. Air Waste Manag. Assoc., 57,
 467 1351–1359, 2007b.

468 469	(20) Liu, Y., Koutrakis, P., Kahn, R., Turquety, S., and Yantosca, R. M.: Estimating
470	fine particulate matter component concentrations and size distributions using
471	satellite-retrieved fractional aerosol optical depth: Part $2 - A$ case study. I
472	AirWaste Manag Assoc 57 1360–1369 2007c
473	7 iii () usio 17 uliug. 7 issoe., 57, 1500 1505, 2007e.
474	(21) Liu, Y., Paciorek, C. J., and Koutrakis, P.: Estimating regional spatial and temporal
475	variability of PM2.5 concentrations using satellite data, meteorology, and land use
476	information, Environ. Health Persp., 117, 886–892, 2009
477	
478	(22) Hoff, R. M. and Christopher, S. A.: Remote sensing of particulate pollution from
479	space: Have we reached the promised land?, J. Air Waste Manag. Assoc., 59, 645-
480	675, 2009, doi: 10.3155/1047-3289.59.6.645;
481	
482	(23) Liu, Y., Park, R. J., Jacob, D. J., Li, Q. B., Kilaru, V., and Sarnat, J. A.: Mapping
483	annual mean ground-level PM2.5 concentrations using Multiangle Imaging
484	Spectroradiometer aerosol optical thickness over the contiguous United States, J.
485	Geophys. Res., 109, D22206, doi:10.1029/2004JD005025, 2004.
486	
487	(24) Xie, Yuanyu; Wang, Yuxuan; Zhang, Kai; Dong, Wenhao; Lv, Baolei and Bai,
488	Yuqi. Daily Estimation of Ground-Level PM2.5 Concentrations over Beijing Using
489	3km Resolution MODIS AOD Environ. Sci. Technol., 2015, 49 (20), pp 12280-
490	12288 DOI: 10.1021/acs.est.5b01413
491	
492	(25) L. A. Remer S. Mattoo, R. C. Levy, and L. A. Munchak. MODIS 3km aerosol
493	product: algorithm and global perspective. Atmos. Meas. Tech., 6, 1829-1844,
494	2013
495	www.atmos-meas-tech.net/6/1829/2013/ doi:10.5194/amt-6-1829-2013
496	
497	(26) Levy, R. C., Remer, L. A., Mattoo, S., Vermote, E. F., and Kaufman, Y. J.: Second-
498	generation operational algorithm: Retrieval of aerosol properties over land from
499	inversion of Moderate Resolution Imaging Spectroradiometer spectral reflectance,
500	J. Geophys. Res., 112, D13211, doi:10.1029/2006JD007811, 2007.
501	

502	(27) Levy, R. C., Remer, L. A., Kleidman, R. G., Mattoo, S., Ichoku, C., Kahn, R., and
503	Eck, T. F.: Global evaluation of the Collection 5 MODIS dark-target aerosol
504	products over land, Atmos. Chem. Phys., 10, 10399-10420, doi:10.5194/acp-10-
505	10399-2010, 2010.
506	
507	(28) Remer, L. A.; Kaufman, Y. J.; Tanre, D.; Mattoo, S.; Chu, D. A.; Martins, J. V.; Li,
508	R. R.; Ichoku, C.; Levy, R. C.; Kleidman, R. G.; et al. Environmental Science &
509	Technology. The MODIS aerosol algorithm, products, and validation. J. Atmos.
510	Sci. 2005, 62 (4), 947–973. doi: 10.1021/acs.est.5b01413;
511	
512	(29) Kees de Hoogh, Harris Héritier, Massimo Stafoggia, Nino Künzli, Itai Kloog,
513	Modelling daily PM2.5 concentrations at high spatio-temporal resolution across
514	Switzerland, In Environmental Pollution, 2017, , ISSN 0269-7491,
515	https://doi.org/10.1016/j.envpol.2017.10.025.
516	
517	(30) Kloog, I.; Nordio, F.; Coull, B. A.; Schwartz, J., Incorporating Local Land Use
518	Regression And Satellite Aerosol Optical Depth In A Hybrid Model Of
519	Spatiotemporal PM2.5 Exposures In The Mid-Atlantic States. Environmental
520	Science & Technology 2012, 46 (21), 11913-11921. dx.doi.org/10.1021/es302673e
521	
522	(31) World Health Organization (WHO) (2006) Air quality guidelines for particulate
523	matter, ozone, nitrogen dioxide and sulfur dioxide—global update 2005 Geneva:
524	WHO Office for Europe.
525	http://www.euro.who.int/data/assets/pdf_file/0005/78638/E90038.pdf
526	
527	(32) Munchak, L. A.; Levy, R. C.; Mattoo, S.; Remer, L. A.; Holben, B. N.; Schafer, J.
528	S.; Hostetler, C. A.; Ferrare, R. A. MODIS 3km aerosol product: applications over
529	land in an urban/suburban region. Atmos. Meas. Tech. 2013, 6 (7), 1747-1759,
530	https://doi.org/10.5194/amt-6-1747-2013
531	
532	(33) Philip K. Hopke , Ying Xie , Taisto Raunemaa , Steven Biegalski , Sheldon
533	Landsberger, Willy Maenhaut , Paulo Artaxo & David Cohen (1997)
534	Characterization of the Gent Stacked Filter Unit PM10 Sampler, Aerosol Science

and Technology, 27:6, 726-735, DOI: 10.1080/02786829708965507 (34) Künzli N, Kutlar Joss MK, Gintowt E (2015) Global standards for global health in a globalized economy! [Editorial]. Int J Public Health 60, 757-759. DOI: 10.1007/s00038-015-0729-0 **TABLES AND FIGURES**

				/	
	Entire peri	iod (25 Sep	tember 200	9 to 21 Oc	ctober 2011)
Variable	\mathbf{N}^{a}	Mean	\mathbf{SD}^{b}	Min	Max
MODIS AOD 3km (unitless)	649	0,29	0,36	0,03	2,19
$PM_{2.5} (\mu g/m^3)$	649	11,36	20,06	1.68	164,41
Average temperature (°C)	649	26,82	1,41	16,24	31,26
Relative humidity (%)	649	84,93	5,80	61,50	98,75
Preciptation (mm)	649	5,06	11,37	0	71,40
		Dry seasor	n (June to N	November))
Variable	Ν	Mean	SD	Min	Max
MODIS AOD 3km (unitless)	323	0,44	0,46	0,03	2,19
$PM_{2.5} (\mu g/m^3)$	323	20,51	25,19	1.68	164,41
Average temperature (°C)	323	27,00	1,68	16,24	31,26
Relative humidity (%)	323	82,18	5,90	61,50	98,75
Preciptation (mm)	323	0	0	0	0
		Rainy sease	on (Decemb	er to May	7)
Variable	Ν	Mean	SD	Min	Max
MODIS AOD 3km (unitless)	326	0,14	0,07	0.03	0,76
$PM_{2.5} (\mu g/m^3)$	326	2,28	2,87	1,68	26,62
Average temperature (°C)	326	26,66	1,07	22,07	29,05
Relative humidity (%)	326	87,65	4,21	73,80	98,00
Preciptation (mm)	326	7,76	14,09	0,10	71,40
		Pear	son correla	ntion	
Variable	AOD	PM2.5	TEMP	RH	PRECIP
MODIS AOD 3km (unitless)	1				
$PM_{2.5}(\mu g/m^3)$	0.5811	1			
Average temperature (°C)	0.245	0.1007	1		
Relative humidity (%)	-0.3957	-0.4455	-0.4411	1	
Preciptation (mm)	-0.101	-0.1454	-0.1638	0.2413	1

Note: a= Number of values observed in days; b=Standard deviation (SD)

549	Table 1: Descriptive statistics of the parameters analysed during the study period
550	(September, 25 th 2009 to October, 21 th 2011).
551	
552	
553	
554	
555	
556	
557	
558	
559	

Prediction models	Ν	df ^a	R²	R^2 adj ^b	RMSE ^c
Model 1	649	3	0.544	0.541	13.59
Model 2	649	18	0.683	0.674	11.47
Model 3	649	21	0.707	0.697	11.04
Model 4	649	27	0.782	0.772	9.58
Model 5	649	15	0.823	0.816	8.60

560 Note: a=degrees of freedom; b= R-squared adjusted; c= Residual standard deviation (RMSE).

561 Model 1 = simple model with linear, quadratic and cubic term of AOD;

562 Model 2 = Model 1 + interactions of AOD, AOD^2 and AOD^3 with linear and quadratic terms in

temperature and relative humidity;

564 Model 3 = Model 2 + interactions between AOD, AOD² and AOD³ and rain;

565 Model 4 = Model 3 + interactions of AOD, AOD^2 and AOD^3 with sine and cosines of date with a period

566 of 365.25 days (without the term for rainy season);

567 Model 5 = Model 4 + lagged relative residual and its square as additional predictor variables; 568

569 **Table 2:** Comparison between prediction models.

570

571

Model 1	Estimate	Std. Error	t value	Pr (> t)
βΟ	113.565	0.5333	21.295	< 2e-16 ***
β1	3.733.880	135.860	27.483	< 2e-16 ***
β2	369.936	135.860	2.723	0.00665 **
β3	319.761	135.860	2.354	0.01889 *
Model 2	Estimate	Std. Error	t value	Pr (> t)
βΟ	1,10E+06	1,89E+05	5.803	1.03e-08 ***
β1	8,13E+07	9,91E+06	8.202	1.33e-15 ***
β2	3,19E+07	8,50E+06	3.751	0.000193 ***
β3	2,32E+06	5,36E+06	0.432	0.665709
β4	8,85E+04	7,41E+04	1.195	0.232703
β5	-1,62E+04	1,26E+04	-1.286	0.198907
β6	-2,44E+03	9,76E+02	-2.503	0.012569 *
β7	4,02E+02	4,75E+02	0.847	0.397177

β8	-4,14E+05	2,03E+05	-2.046	0.041196 *	
β9	2,82E+04	3,92E+04	0.720	0.472077	
β10	9,39E+03	3,01E+03	3.119	0.001895 **	
β11	5,05E+01	1,17E+02	0.431	0.666751	
β12	-1,19E+03	1,17E+03	-1.017	0.309673	
β13	9,79E+04	9,91E+04	0.988	0.323655	
β14	-5,25E+04	2,15E+04	-2.445	0.014744 *	
β15	-3,73E+03	1,53E+03	-2.437	0.015070 *	
β16	9,53E+01	7,64E+01	1.247	0.212722	
β17	1,34E+03	5,54E+02	2.420	0.015795 *	
Model 3	Estimate	Std. Error	t value	Pr(> t)	
β0	1,09E+06	1,85E+05	5.870	7.05e-09 ***	
β1	8,22E+07	9,58E+06	8.584	< 2e-16 ***	
β2	3,36E+07	8,29E+06	4.057	5.60e-05 ***	
β3	2,95E+06	5,21E+06	0.566	0.571754	
β4	8,48E+04	7,17E+04	1.183	0.237140	
β5	-1,68E+04	1,22E+04	-1.379	0.168533	
β6	-2,47E+03	9,43E+02	-2.616	0.009109 **	
β7	5,13E+02	4,60E+02	1.116	0.264968	
β8	-4,16E+05	1,96E+05	-2.123	0.034177 *	
β9	4,21E+04	3,87E+04	1.088	0.276841	
β10	9,67E+03	2,91E+03	3.325	0.000935 ***	
β11	-1,90E+01	1,19E+02	-0.159	0.873546	
β12	-1,46E+03	1,14E+03	-1.285	0.199287	
β13	8,94E+04	9,59E+04	0.932	0.351465	
β14	-5,79E+04	2,12E+04	-2.734	0.006433 **	
β15	-3,69E+03	1,48E+03	-2.498	0.012757 *	
β16	1,20E+02	7,65E+01	1.572	0.116357	
β17	1,46E+03	5,39E+02	2.708	0.006950 **	
β18	-1,01E+05	2,00E+04	-5.063	5.44e-07 ***	
β19	3,62E+05	1,18E+05	3.063	0.002283 **	
β20	-2,58E+05	1,33E+05	-1.935	0.053453.	
Model 4	Estimate	Std. Error	t value	Pr (> t)	
βΟ	5,55E+05	1,68E+05	3.305	0.001003 **	
β1	3,85E+07	1,04E+07	3.692	0.000242 ***	
β2	4,62E+06	9,84E+06	0.469	0.638882	
β3	-6,61E+06	5,32E+06	-1.245	0.213777	
β4	1,04E+05	6,25E+04	1.659	0.097687.	
β5	1,28E+02	1,08E+04	0.012	0.990548	
β6	-1,88E+03	8,21E+02	-2.291	0.022301 *	
β7	-1,21E+02	4,08E+02	-0.296	0.767175	
β8	-3,37E+05	1,72E+05	-1.956	0.050898.	
β9	-5,94E+04	4,05E+04	-1.467	0.142785	
β10	4,84E+03	2,56E+03	1.893	0.058828.	
β11	2,10E+02	1,19E+02	1.773	0.076650.	
β12	9,88E+02	1,11E+03	0.894	0.371821	
β13					
	9,74E+04	8,57E+04	1.136	0.256252	

β15	-1,46E+03	1,30E+03	-1.129	0.259177	
β16	-8,98E+01	8,23E+01	-1.091	0.275639	
β17	-2,22E+02	5,79E+02	-0.383	0.701887	
β18	-1,24E+04	2,54E+04	-0.488	0.625602	
β19	7,89E+04	1,19E+05	0.664	0.506904	
β20	-9,50E+04	1,24E+05	-0.767	0.443244	
β21	6,99E+04	1,49E+04	4.681	3.51e-06 ***	
β22	-3,71E+04	7,47E+03	-4.966	8.83e-07 ***	6
β23	-2,64E+05	5,20E+04	-5.068	5.31e-07 ***	
β24	4,90E+04	2,17E+04	2.257	0.024378 *	
β25	1,08E+05	3,06E+04	3.545	0.000423 ***	
β26	-3,07E+04	1,37E+04	-2.244	0.025199 *	
Model 5	Estimate	Std. Error	t value	Pr (> t)	
β0	7,46E+03	1,19E+03	6.266	6.92e-10 ***	
β1	1,18E+06	1,62E+06	0.732	0.464505	
β2	4,13E+06	3,97E+06	1.041	0.298197	
β3	-1,48E+06	1,97E+06	-0.754	0.451224	
β4	5,84E+04	6,22E+04	0.938	0.348512	
β5	-4,32E+04	2,33E+04	-1.858	0.063575.	
β6	-1,53E+03	7,45E+02	-2.049	0.040855 *	
β7	2,09E+02	8,55E+01	2.443	0.014849 *	
β8	2,10E+02	4,51E+02	0.465	0.641789	
β9	-3,03E+05	1,66E+05	-1.833	0.067247.	
β10	5,14E+01	5,64E+04	0.001	0.999274	
β11	4,05E+03	2,31E+03	1.750	0.080635 .	
β12	-1,70E+02	2,04E+02	-0.833	0.404931	
β13	1,11E+03	1,13E+03	0.981	0.327124	
β14	9,58E+04	8,07E+04	1.187	0.235520	
β15	2,21E+03	2,94E+04	0.075	0.939958	
β16	-1,14E+03	1,17E+03	-0.975	0.329859	
β17	5,66E+01	1,06E+02	0.533	0.594386	
β18 0.10	-4,24E+02	5,72E+02	-0.741	0.459242	
β19	7,01E+04	7,96E+03	8.812	< 2e-16 ***	
β20	-3,67E+04	5,74E+03	-6.389	3.27e-10 ***	
β21	-2,55E+05	3,53E+04	-7.223	1.49e-12 ***	
β22	5,66E+04	1,77E+04	3.195	0.001470 **	
β23	1,18E+05	2,26E+04	5.234	2.27e-07 ***	
β24	-4,10E+04	1,17E+04	-3.504	0.000491 ***	
β25	4,55E+02	3,75E+01	12.129	< 2e-16 ***	
Model 6	Estimate	Std. Error	t value	$\Pr(> t)$	
βU	2,35E+03	9,58E+01	24.568	< 2e-16 ***	
β1 02	-1,12E+04	5,55E+03	-2.019	0.0439 *	
β2 02	2,02E+04	2,16E+03	9.358	< 2e-16 ***	
β3 04	-9,94E+03	9,56E+02	-10.405	< 2e-16 ***	
β4 05	1,71E+02	3,39E+02	0.503	0.6154	
β5 Ac	5,82E+01	7,62E+01	0.764	0.4454	
po	-3,13E+00	6,32E+00	-0.495	0.6211	
p/	-4,23E-01	4,83E-01	-0.875	0.3817	

β8	8,42E+03	9,67E+02	8.708	< 2e-16 ***
β9	-6,59E+03	4,36E+02	-15.117	< 2e-16 ***
β10	-2,08E+04	1,97E+03	-10.561	< 2e-16 ***
β11	1,01E+04	8,26E+02	12.217	< 2e-16 ***
β12	1,00E+04	8,98E+02	11.130	< 2e-16 ***
β13	-4,31E+03	3,51E+02	-12.275	< 2e-16 ***
β14	5,57E+02	4,34E+01	12.841	< 2e-16 ***
β15	-8,85E+01	1,11E+01	-8.012	5.45e-15 ***

Note: Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

574 **Table 3:** Description of parameters, standard error and p-value for each prediction





Figure 2: (A) Comparisons between $PM_{2.5}$ measured and $PM_{2.5}$ predictions ($\mu g/m^3$) across time (25 September 2009 to 21 October 2011). (B) Partial autocorrelation plot of residuals before introducing the lagged relative residual and its square as additional predictor variables into the model (C) Partial residual autocorrelation plot after after adding the the two variables to the model.



Figure 3: Comparisons between the measured and predicted $PM_{2.5}(\mu g/m^3)$ for (A) Model 2 and (B) non-linear prediction (Final model). The red line represents the regression line.

- 663
- 664
- 665
- 666
- 667
- 668



	ACCEPTED MANUSCRIPT
704	
705	
706	
707	
708	
709	
710	
711	Figure 4: Spatial distribution of PM _{2.5} predicted concentration over the basin during
712	different seasons. Forest fires occurs between the months of September, October and
713	November. (A): PM _{2.5} averages predicted concentration interpolated to all Brazilian
714	states; (B): Rondônia State including the study area of Porto Velho.

Highlights

- Non-linear model was applied for predicting $PM_{2.5}$ from AOD with a good performance;
- The model can be applied to other sites if site-specific data are available;
- The lagged relative residual it is cautious strategy to further improve the model;
- It was the first Brazilian study predicting PM_{2.5} from high resolution AOD data;